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Research Article

Treatment f organophenolic and organoaromatics from textile wastewater using NiCo₂O₄ doped Bi₂O₂CO₃ nanocomposite

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Abstract

The textile wastewaters could not be treated effectively with conventional treatment processes due to high polyphenol and aromatic compounds and colour content. In this study, by doping of NiCo₂O₄ to Bi₂O₂CO₃ the generated NiCo₂O₄ to Bi₂O₂CO₃ nanocomposite was used for the photocatalytic oxidation of COD components (COD_{total} COD_{dissolved} COD_{dissolved} COD_{inert}), color, organophenols and organoaromatic compounds from a textile industry wastewaters (TW) at different operational conditions such as, at different photooxidation times (5 min, 15 min, 30 min, 60 min, 80 min and 100 min), at different NiCo₂O₄ ratios (0.5wt% , 1wt%, 1.5wt%, 2wt%), at different NiCo₂O₄ / Bi₂O₂CO₃ nanocomposite concentrations (1, 5, 15, 30 and 45 mg/L), under 10, 30, 50 and 100 W solar irradiations, respectively. The maximum COD_{total} COD_{inert}, total flavonols, total aromatic amines (TAAs) and color photooxidation yields were 99%, 92%, 91%, 98% and 99% respectively, under the optimized conditions, at 30 mg/L Ni/BiO nanocomposite with a Ni mass ratio of 1.5 wt% under 50 W UV (ultraviolet) light, after 60 min photooxidation time, at 25°C. The photooxidation yields of kaempferol (KPL), quercetin (QEN), patuletin (PTN), rhamnetin (RMN) and rhamnazin (RHAZ) from flavonols and 2-methoxy-5-methylaniline (MMA), 2,4-diaminoanisole(DAA); 4,40-diamino diphenyl ether (DDE), o-aminoazotoluene (OAAT), and 4-aminoazobenzol (AAB) from polyaromatic amines were > 82%. The pollutants of textile industry wastewater were effectively degraded with Ni doped BiO nanocomposite.

Keywords: Flavonols; Nickel cobaltite NiCo₂O₄ nanocomposite; bismuth subcarbonate ($Bi_2O_2CO_3$) nanocomposite; Photooxidation; Polyaromatic amines; Ultraviolet (UV) light irradiation.

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Introduction

Textile industry is one of those industries that consume large amounts of water in the manufacturing process [1] and, also, discharge great amounts of effluents with synthetic dyes to the environment causing public concern and legislation problems. Synthetic dyes

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that make up the majority (60-70%) of the dyes applied in textile processing industries [2] are considered to be serious health risk factors. Apart from the aesthetic deterioration of water bodies, many colorants and their breakdown products are toxic to aquatic life [3] and can cause harmful effects to humans [4, 5]. Several physico-chemical and biological methods for dye removal from wastewater have been investigated [6-8] and seem that each technique faces the facts of technical and economical limitations [7]. The traditional physical, chemical and biologic means of wastewater treatment often have little degradation effect on this kind of pollutants. On the contrary, the technology of nanoparticulate photodegradation has been proved to be effective to them. Compared with the other conventional wastewater treatment means, this technology has such advantages as: (1) wide application, especially to the molecule structure-complexed contaminants which cannot be easily degraded by the traditional methods; (2) the nanoparticles itself have no toxicity to the health of our human livings and (3) it demonstrates a strong destructive power to the pollutants and can mineralize the pollutants into carbondipxide (CO₂) and water (H₂O) [9].

 ${\rm Bi_2O_2CO_3}$ has gained much attention due to its promising photocatalytic activity for wastewater treatment [10-12]. Although

Bi₂O₂CO₂ has been widely studied in the photocatalytic degradation of wastewater, little attention has been poured to investigate the microwave catalytic performance of Bi₂O₂CO₂ for microwave catalytic oxidation degradation of wastewater, up to now. At the same time, the magnetic NiCo₂O₄ has intriguing advantages, such as excellent microwave absorption performance, low cost, magnetically separable property, and high stability [13]. To the best of our knowledge, NiCo₂O₄-Bi₂O₂CO₂ composite as microwave catalyst for degradation o more semiconductor photocatalysts have been found to be capable of photocatalytic degradation of organic macromolecular contaminants in wastewater [14, 15]. Therefore, photocatalytic degradation has become the most environmentally friendly, energy-saving, and efficient water pollution treatment method. In view of the fact that the traditional photocatalysts (such as TiO,) have large band gap energy and low response to visible light, their application is greatly limited. Among these miconducting photocatalysts, bismuth molybdate (Bi₂MoO₆) as a ternary oxide compound of Aurivillius phase becomes one of the promising materials. This is because it has a unique layered structure sandwiched between the perovskite octahedral (MoO₄)₂sheets and bismuth oxide layers of (Bi2O2)2+ [16-18]. Its dielectric property, ion conductivity, and catalytic performance have obvious advantages in bismuth-based semiconductors [19, 20]. Nevertheless, the light absorption property of the pure Bi₂MoO₆ primarily appears in the ultraviolet light region, which is only a small part of the solar spectra. Meanwhile, it presents a high recombination rate of electronhole pairs in the process of photocatalytic reaction [21]. Therefore, researchers have improved the performance of Bi₂MoO₆ by means of morphology controlling, semiconductor compounding, and doping modification [22]. Among these measures, doping has proven to be an effective method to ameliorate the surface properties of photocatalysts and enhance photocatalytic performance.

It was reported that carbon-doped Bi₂MoO₆ exhibited significantly enhanced and stable photocatalytic properties compared with Bi₂MoO₆ [23], which carbon replaced the O₂⁻ anion in the lattice of Bi₂MoO₆, resulting in lattice expansion and grain diameter reduction, enhancement of specific surface area [24]. prepared Graphene-Bi₂MoO₆ (G-Bi₂MoO₆) hybridphotocatalysts by a simple one-step process, and an increase in photocatalytic activity was observed for G-Bi₂MoO₆ hybrids compared with pure Bi₂MoO₆ under visible light. Xing et al., (2017) reported the photocatalytic activity of 0.5% Pd-3C/BMO was robustly enhanced about 5-fold for Rhodamine B (RhB) degradation within 40 min under UV + visible light irradiation and 29-fold for O-phenylphenol (OPP) degradation within 120 min under visible light irradiation in comparison with pristine Bi₂MoO₄, respectively. [25] prepared a B-doped Bi₂MoO₆ photocatalyst with hydrothermal method by using HBO3 as a dopant source. It was found that B-doping increases the amount of Bi5+ and oxygen vacancies, so that the visible light absorption of catalyst is stronger, and the band gap energy is lower, which significantly improves the photocatalytic activity of Bi₂MoO₆ [26] successfully synthesized sulfur-doped copper-cobalt bimetal oxide by coprecipitation method, which significantly improved the catalytic performance and stability of the catalyst. [27] fabricated Bi, MoO, surface co-doped with Ni²⁺ and Ti4+ ions through an incipient-wetness impregnation technology and calcination method, with the results suggesting Ni^{2+} and Ti^{4+} codoping increases visible-light absorption by Bi_2MoO_6 and promotes the separation of photogenerated charge carriers. Density functional theory calculations and systematical characterization results revealed that Biself-doping could not only promote the separation and transfer of photo generated electron-hole pairs of Bi_2MoO_6 but also alter the position of valence and conduction band without changing its preferential crystal orientations, morphology, visible light absorption, as well as band gap energy [28, 29] synthesized pure and various contents of Ce^{3+} doped Bi_2MoO_6 nano structures by a facile hydrothermal method. The 0.5%Ce³⁺ doped Bi_2MOO_6 exhibits best photocatalytic activity of 96.6% within 20 min for RhB removal.

The photocatalytic performance of NiCo₂O₄-doped Bi₂MoO₆ nanoparticles has not been investigated extensively for the removals of aromatics and polyphenols from a textile industry. In this work, the phsicochemical properties of NiCo₂O₄ doped Bi₂O₂CO₂ nanocomposite was investigated using microscope (SEM), transmission electron microscopy (TEM), Fourier transform infrared spectroscopy (FT-IR), X-ray photoelectron spectroscopy (XPS), photoluminescence spectra (PL), N₂ adsorption-desorption, elemental mapping, Raman and diffused reflectances pectra (DRS) analysis. The photocatalytic oxidation of pollutant parameters [COD components (COD_{total}, COD_{dissolved}, COD_{inert}), flavonols (kaempferol, quercetin, patuletin, rhamnetin and rhamnazin), polyaromatic amines (2-methoxy-5-methylaniline, 2,4-diaminoanisole, 4,40-diamino diphenyl ether, o-aminoazotoluene, and 4-aminoazobenzol) and color] from the TW at different operational conditions such as, at increasing photooxidation times (5 min, 15 min, 30 min, 60 min, 80 min and 100 min), at diferent Ni mass ratios (0.5wt%, 1wt%, 1.5wt%, 2wt%), at different Ni-BiO photocatalyst concentrations (1, 5, 15, 30 and 45 mg/L), at different pH ranges (4, 6, 8, 10) under 10, 30, 50 and 100 W UV light irradiations, respectively, were investigated .

Materials and methods

Raw wastewater

The characterization of raw TW was given in Table 1.

Chemical structure of flavonols and poliaromatics present in the TW

The structure of flavonols in the TW was shown in **Figure 1**. The structure of polyaromatics in the TW was given **Figure 2**.

Preparation of photocatalysts

Ni-doped BiO nano particles were prepared by co-precipitation method using nickel nitrate hexahydrate [Ni(NO₃)₂.6H₂O] (Analytical grade, Merck) and Bismuth nitrate hexahydrate [Bi(NO₃)₂.6H₂O] (Sigma, Aldrich) as the precursors of nickel and bismuth, respectively. Ni(NO₃)₂.6H₂O and sodium carbonate anhydrous (Na₂CO₃) were dissolved separately in double distilled H₂O to obtain 0.5 mol/L solutions. Nickel nitrate solution (250 mL of 0.5 mol/L) was slowly added into vigorously stirred 250 mL of 0.5 mol/L Na₂CO₃ solution. Nickel nitrate in the required stoichiometry was slowly added into the above solution and a white precipitate was obtained. The precipitate

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aParameters	Minimum	Medium	Maximum
pН	5.00 ± 0.18	5.27 ± 0.19	6.00 ± 0.21
DO (mg/L)	1.30 ± 0.05	1.40 ± 0.05	1.50 ± 0.05
ORP (mV)	85.00 ± 2.98	106.00 ± 3.71	128.00 ± 4.48
TSS (mg/L)	285.00 ± 9.98	356.00 ± 12.46	430.00 ± 15.05
TVSS (mg/L)	192.00 ± 6.72	240.00 ± 8.40	290.00 ± 10.15
COD _{total} (mg/L)	931.70 ± 32.61	1164.60 ± 40.76	1409.20 ± 49.32
COD _{dissolved} (mg/L)	770.40 ± 26.96	962.99 ± 33.71	1165.22 ± 40.78
TOC (mg/L)	462.40 ± 16.18	578.00 ± 20.23	700.00 ± 24.50
$BOD_5 (mg/L)$	251.50 ± 8.80	314.36 ± 11.00	380.38 ± 13.31
BOD ₅ /COD _{dis}	0.26 ± 0.01	0.33 ± 0.012	0.40 ± 0.014
Total N (mg/L)	24.80 ± 0.87	31.00 ± 1.09	37.51 ± 1.31
NH ₄ -N (mg/L)	1.76 ± 0.06	2.20 ± 0.08	2.66 ± 0.09
NO ₃ -N (mg/L)	8.00 ± 0.28	10.00 ± 0.35	12.10 ± 0.42
NO ₂ -N (mg/L)	0.13 ± 0.05	0.16 ± 0.06	0.19 ± 0.07
Total P (mg/L)	8.80 ± 0.31	11.00 ± 0.39	13.30 ± 0.47
PO ₄ -P (mg/L)	6.40 ± 0.22	8.00 ± 0.28	9.68 ± 0.34
SO ₄ ⁻² (mg/L)	1248.00 ± 43.70	1560.00 ± 54.60	1888.00 ± 66.10
Color (1/m)	70.90 ± 2.48	88.56 ± 3.10	107.20 ± 3.75
Flavonols (mg/L)	30.9 ± 1.08	38.6 ± 1.35	46.1 ± 1.61
Flavonols			
Kaempferol	4.2 ± 0.20	5.7 ± 0.2	7.2 ± 0.3
Quercetin	7.3 ± 0.26	9.2 ± 0.32	11.1 ± 0.4
Patuletin	8.3 ± 0.30	10.3 ± 0.36	12.2 ± 0.43
Rhamnetin	6.0 ± 0.21	7.2 ± 0.25	8.4 ± 0.3
Rhamnazin	5.1 ± 0.18	6.15 ± 0.22	7.2 ± 0.25
TAAs (mg benzidine/L)	891.84 ± 31.21	1038 ± 36.33	1183.8 ± 41.43
Polyaromatics			
2-methoxy-5-methylaniline	128.5 ± 4.5	134.6 ± 4.71	140.6 ± 4.92
2,4-diaminoanisole	250.2 ± 8.76	275.8 ± 9.7	301.3 ± 10.6
4,40-diamino diphenyl ether	146.54 ± 5.13	156.0 ± 5.5	165.4 ± 5.8
o-aminoazotoluene	265.4 ± 9.3	293.6 ± 10.3	321.7 ± 11.3
4-aminoazobenzol	101.2 ± 3.54	178 ± 6.23	254.8 + 8.92

Table 1: Characterization values of TW at pH=5.7 (n=3, mean values \pm SD). (SD: standard deviation; n: the repeat number of experiments in this study).

was filtered, repeatedly rinsed with distilled H_2O and then washed twice with ethanol. The resultant solid product was dried at 100°C for 12 h and calcined at 300°C for 2 h. BiO particles were also prepared by the same procedure without the addition of nickel nitrate solution. The doping Ni mass ratios of Bismuth are expressed as wt%.

X-Ray diffraction (XRD) analysis

XRD patterns of the samples are going to carry out using a D/Max-2400Rigaku X-ray powder diffractometer operated in the reflection mode with Cu Ka ($\lambda = 0.15418$ nm) radiation through scan angle (2 θ) from 20° to 80°.

Scanning electron microscopy (SEM) analysis

The morphological structures of the Ni-BiO nanocomposites before photocatalytic degradation with UV light irradiations and after photocatalytic degradation with UV by means of a SEM.

Fourier transform infrared spectroscopy (FTIR) analysis

The FTIR spectra of Ni, BiO and Ni-BiO samples were measured with FTIR spectroscopy measurements.



Figure 1: Chemical structure of flavonoids in the TW.



Figure 2: Chemical structure of polyaromatics in the TW

Photocatalytic degradation reactor

A 2 L cylinder kuvars glass reactor was used for the photodegradation experiments in the TW under different UV powers, at different operational conditions. 1000 mL TW was filled for experimental studies and the photocatalyst were added to the cylinder glass reactor. The photocatalytic reaction was operated with constant stirring during the photocatalytic degradation process. 10 mL of the reacting solution were sampled and centrifugated (at 10000 rpm) at different time intervals.

Used chemicals

Ni(NO₃)₂.6H₂O (Analytical Grade, Merck, Germany) and Bi(NO₃)₃.6H₂O (Analytical grade, Merck, Germany) were used as nickel and bismuth sources, respectively. Na₂CO₃ was purchased from Merck (Analytical grade). Helium, He(g) (GC grade, 99.98%) and nitrogen, N₂(g) (GC grade, 99.98%) was purchased from Linde, (Germany). Kaempferol (99%), quercetin (99%), patuletin (99%), rhamnetin (99%), rhamnazin (99%), 2-methoxy-5-methylaniline (99%), 2,4-diaminoanisole (99%), 4,40-diamino diphenyl-ether (99%), o-aminoazotoluene (99%), 4-aminoazobenzol (99%) were purchased from Aldrich, (Germany).

Analytical methods

pH, T(°C), ORP, DO, BOD_5 , COD_{total} , $COD_{dissolved}$, total suspended solids (TSS), Total-N, NH₃-N, NO₃-N, NO₂-N, Total-P and PO₄-P measurements were monitored following the Standard Methods 2310, 2320, 2550, 2580, 4500-O, 5210 B, 5220 D, 2540 D, 4500-N, 4500-NH₃, 4500-NO₃, 4500-NO₂ and 4500-P [30]. Inert COD was measured according to glucose comparison method [31]. The samples were analyzed by high pressure liquid chromatography (HPLC) with photodiode array and mass spectrometric detection using an Agilent 1100 high performance liquid chromatography system consisting of an automatic injector, a gradient pump, a Hewlett-Packard series 1100 photodiode array detector, and an Agilent series 1100 VL online atmospheric pressure ionization electrospray ionization mass spectrometer to detect flavonols namely kaempferol, quercetin, patuletin, rhamnetin, rhamnazin and polyaromatics namely, 2-methoxy-5-methylaniline, 2,4-diaminoanisole, 4,40-diamino diphenyl-ether, o-aminoazotoluene, 4-aminoazobenzol, respectively. All the metabolites were measured in the same HPLC by mass spectrometric detections. Operation of the system and data analysis were done using ChemStation software, and detection was generally done in the negative ion $[M - H]^-$ mode, which gave less complex spectra, although the positive ion mode was sometimes used to reveal fragmentation patterns-especially patterns of sugar attachment. Separation of flavonol components was made on a Vydac C18 reversed phase column (2.1 µm dia. × 250 mm long; 5-µm particle size). Columns were eluted with acetonitrile-water gradients containing 0.1% formic acid in both solvents. The quality of the raw (untreated) and photooxidated wastewater were determined by measuring the absorbances of the supernatans at wavelengths varying between 200 nm, 250 nm, 300 nm, 350 nm and 540 nm using an Aquamate Termoelectron Corporation UV-vis spectrophotometer.

Measurement of photonic efficiency (l,) of Ni doped BiO

The relative photonic efficiency of the catalyst is obtained by comparing the photonic efficiency of Ni-doped BiO with that of the standard photocatalyst (BiO). In order to evaluate l_{p} a solution of 1-Methylcyclopropene-MCP (40 mg/L) with a pH of 10 was irradiated with 100 mg of BiO and Ni-doped BiO for 60 min. From the degradation results, Ir was calculated as follows (Eq. 1).

$$I_r = \frac{\text{initial rate of MCP degradation on Ni-doped BiO}}{\text{initial rate of MCP degradation on pure BiO}}$$
(1)

Operational conditions

Under 10-30-50 and 100 W UV light powers the photocatalytic oxidation of the pollutant parameters in the TW at different operational conditions such as at increasing Ni mass ratios in the Ni-BiO nanocomposite (0.5wt%, 1wt%, 1.5wt%, 2wt%), at increasing photooxidation times (5 min, 15 min, 30, 60 min, 80 min and 100 min), at different Ni-BiO photocatalyst concentrations (1, 5, 15, 30 and 45 mg/L), under acidic, neutral and basic conditions, respectively.

All the experiments were carried out following the batch-wise procedure. All experiments were carried out three times and the results were given as the means of triplicate sampling with standard deviation (SD) values.

Results and analysis

XRD Analysis results

The powder XRD patterns of BiO and Ni-doped BiO with different lanthanum mass ratios are shown in **Figure 3**. The XRD patterns of all the Ni-doped BiO catalysts are almost similar to that of BiO, suggesting that there is no change in the crystal structure upon Ni loading. This also indicates that Ni⁺² is uniformly dispersed on BiO nanoparticles in the form of small Ni₂O₂ cluster. However the Nidoped samples have a wider and lower intense diffraction peaks than pure BiO. Moreover, the XRD peaks of Ni-doped BiO continuously get broader with increasing the Ni loading up to a mass ratio of 2%wt.

SEM Analysis results

The morphology of nanocomposite particles is analyzed by SEM.



Figure 3 XRD patterns of BiO and Ni doped BiO (a) pure BiO, (b) 2 wt% Ni doped BiO, (c) 0.5wt% Ni doped BiO, (d) 1.0wt% Ni doped BiO, and (e) 1.5wt% Ni doped BiO.

Figure 4 shows that the nanocomposite material is partly composed of clusters containing composite nanoparticles adhering to each other with a mean size of around 20-80 nm before photooxidation process (**Figure 4a**) while the size increased to 24-86 nm after photooxidation (**Figure 4b**) with intermediates and remaining not photodegraded pollutants.

FTIR Analysis results

Figure 5 shows the FTIR spectrum of BiO and Ni-doped BiO, BiO powder synthesized under laboratory conditions. The peak between 400 and 700 cm⁻¹ give the information of Bi–O and Ni–Bi–O on the FTIR spectra. The peak at 437–455 cm⁻¹ give the information about stretching vibration of crystalline hexagonal zinc oxide (Bi– O stretching, vibration) and the peaks from 902 to 1020 cm⁻¹ are attributed to the bond between lanthanum and oxygen (Ni–O). The broad peak between 3400 to 3900 cm⁻¹ indicate the OH groups, due to the H₂O which indicates the existence of atmospheric H₂O adsorbed on the surface of nanocrystalline powder. An absorption band and a peak have been observed at 2350 cm⁻¹, respectively, which arises from the absorption of atmospheric CO₂ on the metal cations.

Results and Discussions

Effect of increasing Ni-BiO nanocomposite concentrations on the removals of TW pollutants

The effects of increasing Ni-BiO nanocomposite concentrations (1 mg/L, 5 mg/L, 15 mg/L, 30 mg/L and 45 mg/L), on the photocatalytic



Figure 4 SEM micrographs of pure and nickel modified BiO, (a) pure BiO at 25°C, (b) Ni doped BiO at 25°C.



Figure 5: FTIR Spectra of pure BiO and Ni-doped BiO nanoparticles, with different concentration of dopant

oxidation of polutant parameters in the TW was investigated. The preliminary studies showed that the maximum removal of COD with 20 mg/L Ni-BiO nanocomposite was 89% with 70 min photooxidation time at pH=7.8 with 40 W UV power (Data not shown). Based on these yields the operational conditions for photocatalytic time were choosen as 60 min at a power of 50 W and at a pH of 8. The maximum photocatalytic oxidation removals for all pollutants in the TW were observed at 30 mg/L Ni-BiO nanocomposite concentrations, at pH=8.0, after 60 min photooxidation time and at 25°C at a power of 50 W (Figure 6). Removal efficiencies slightly decreased at 45 mg/L Ni-BiO nanocomposite concentration, because over load of surface area of Ni-BiO nanocomposites (Figure 6). This limiting the power of UV irradiation. Lower photo-removal efficiencies was measured for 1, 5, and 15 mg/L Ni-BiO concentrations due to low surface areas in the nanocomposite. On the contrarily, the surface area is high at 30 mg/L Ni-BiO nanocomposite concentrations. Therefore, the maximum photodegradation yield was observed in this nanocomposite concentration. The $\text{COD}_{\text{total}}$, $\text{COD}_{\text{inert}}$, total flavonols, total aromatic amines and color removals increased linearly as the Ni-BiO nanocomposite concentrations were increased from 1 mg/L



■ Ni-BiO=1 mg/L
^{III} Ni-BiO=5 mg/L
^{III} Ni-BiO=15 mg/L
^{III} Ni-BiO=30 mg/L

Figure 6: Removal efficiencies of COD_{iotal} COD $_{iotal}$ total flavonols and TAAs at Ni-BiO=1 mg/L, BiO=5 mg/L, BiO=15 mg/L and BiO=30 mg/L.

up to 5 mg/L, to 15 mg/L, and up to 30 mg/L, respectively (Table 2 and Figure 6). Furher increase of nanocomposite concentration to 45 mg/L affect negatively the all the pollutant yields. The reason for this is the optimum amount of catalyst increases the number of active sites on the photocatalyst surface, which in turn increase the number of OH $^{\bullet}$ and superoxide radicals (O₂ $^{-\bullet}$) to degrade pollutant parameters (COD components, flavonols, polyaromatics, color). When the concentration of the catalyst increases above the optimum value, the degradation decreases due to the interception of the light by the suspension [32]. reported that as the excess catalyst (turbidity) prevent the illumination of light, OH[•], a primary oxidant in the photocatalytic system decreased and the efficiency of the degradation reduced accordingly. Furthermore, the increase in catalyst concentration beyond the optimum may result in the agglomeration of catalyst particles; hence, the part of the catalyst surface becomes unavailable for photon absorption, and thereby, photocatalytic oxidation efficiency decreases [33]. Maximum COD_{total}, COD_{inert}, total flavonols, TAAs and color removal efficiencies were obtained after 60 min photooxidation process with yields of 99%, 92%, 91%, 98% and 99%, respectively, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W power and at 25°C (Figure 6). Flavonols such as kaempferol, quercetin, patuletin, rhamnetin, rhamnazin removal efficiencies were 87%, 88%, 90%, 87% and 85% respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration and at 25°C temperature (Table 2). Polyaromatic amines such as, 2-methoxy-5-methylaniline, 2,4-diaminoanisole, 4,40-diamino diphenyl ether, o-aminoazotoluene, 4-aminoazobenzol removal efficiencies after photooxidation process were 93%, 95%, 87%, 84% and 82%, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (Table 2).

Kaempferol metabolies such as, 3-O-[2-O, 6-O-bis (α-L- rhamnosyl)-(β -D-glucosyl] quercetin, 3-O-[6-O-(α -L -rhamnosyl)-(ß-D - glucosyl] quercetin, 3-O-{2-O-[6-O-(phydroxy-trans-cinnamoyl)-{)-, β -D -glucosyl]- \dot{a} -L- rhamnosyl} kaempferol decreased from 5.7 mg/L to 0.86 mg/L, from 5.7 mg/L to 1.08 mg/L, from 5.7 mg/L to 1.25 mg/L, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (Table 3). Quercetin metabolites such as, 3-O-[6-O-(α -L -rhamnosyl)-)-(β -D-glucosyl] quercetin, 3-O-{2-O-[6-O-(p-hydroxy-trans-cinnamoyl)-(ß-D -glucosyl]-- á-L- rhamnosyl} quercetin reduced from 9.2 mg/L to 1.28 mg/L, from 9.2 mg/L to 2.30 mg/L, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (Table 3). Patuletin metabolites such as, (E)-ascladiol, (Z)-ascladiol dropped off from 10.3 mg/L to 1.55 mg/L, from 10.3 mg/L to 1.85, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (Table 3). Rhamnetin metabolites such as, methyl quercetin, tetrahydroxy-7-methoxyflavone decreased from 7.2 mg/L to 1.15 mg/L, from 7.2 mg/L to 1.44 mg/L, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (Table 3). Rhamnazin metabolites

 Table 2 Effect of increasing Ni-BiO nanocomposite concentrations on the TW during photooxidation process after 60 min, at 50 W UV irradiation, at pH=8.0, at 25°C.

	Removal efficiencies (%)						
Parameters		Ni-BiO co	oncentratio	ns (mg/L)			
1 arameters	1	5	15	30	45		
	mg/L	mg/L	mg/L	mg/L	mg/L		
COD _{total}	51	65	84	99	79		
COD	45	63	78	92	76		
COD _{dissolved}	50	64	82	98	80		
Color	62	69	85	99	83		
Total flavonols	40 58 79 91			91	72		
Flavonols							
Kaempferol	35	57	72	87	65		
Quercetin	36	61	73	88	67		
Patuletin	37	62	79	90	74		
Rhamnetin	38	56	72	87	64		
Rhamnazin	34	53	71	85	66		
TAAs	58	75	81	98	77		
Polyaromatics							
2-methoxy-5-methylaniline	55	66	83	93	79		
2,4-diaminoanisole	54	71	79	95	73		
4,40-diamino diphenyl ether	52	58	68	87	63		
o-aminoazotoluene	49	65	75	84	72		
4-aminoazobenzol	47	62	76	82	70		

Table 3: The metabolites of flavonols in the TW

Flavonoids	Flavonoids metabolites	Influent concentrations (mg/L)	Effluent Concentrations (mg/L)	Removal efficiencies (%)
FlavonoidsFlavonoids metabolitesSaempferol3-O-[2-O,6-O-bis (α -L-rhamnosyl)-(β -D-glucosyl]- quercetin3-O-[6-O-(α -L -rhamnosyl)-(β -D - glucos quercetin3-O-[2-O-[6-O-(p -hydroxy trans-cinnamoyl)-{})-, β -D -glucosyl]- \dot{a} -L-rhamnosyl] varas-cinnamoyl)-{})-(β -D glucosyl] quercetin3-O-[2-O-[6-O-(α -L -rhamnosyl)-)-(β -D -glucosyl] quercetin3-O-[2-O-[6-O-(α -L -rhamnosyl)-)-(β -D -glucosyl] quercetin3-O-[2-O-[6-O-(α -L -rhamnosyl)-)-(β -D -glucosyl] \dot{a} -L-rhamnosy quercetinPatuletin(E)-ascladiol (Z)-ascladiol(Z)-ascladiol (Z)-ascladiolRhamnetinMethyl quercetin Tetrahydroxy-7- methoxyflavoneRhamnazin 	5.7	0.86	85	
	$3-O-[6-O-(\alpha-L)-rhamnosyl)-(\beta-D-glucosyl]$ quercetin	5.7	1.08	81
	3-O-{2-O-[6-O-(<i>p</i> -hydroxy- <i>trans</i> -cinnamoyl)-{)-, β-D -glucosyl]- <i>ά</i> -L- rhamnosyl} kaempferol	5.7	1.25	78
Quercetin 3-O-[6-O-(α-L -rhamnosyl)-)-(β-D- glucosyl] quercetin		9.2	1.28	86
	3-O-{2-O-[6-O-(<i>p</i> -hydroxy- <i>trans</i> -cinnamoyl)-(<i>f</i> -D -glucosyl] <i>á</i> -L- rhamnosyl} quercetin	9.2	2.30	75
Patuletin	(E)-ascladiol	10.3	1.55	85
	(Z)-ascladiol	10.3	1.85	82
Rhamnetin	Methyl quercetin	7.2	1.15	84
	Tetrahydroxy-7- methoxyflavone	7.2	1.44	80
Rhamnazin	Rhamnazin-3-0-β-D -glucopyranosyl-(l →5)- α-L- arabinofuranoside	6.15	1.42	77
	Rhamnazin-3-O-β-D- glucopyranosyl-(1->5)-[β- D-apiofuranosyl-(-1>2)]-α -L-arabinofuranoside	6.15	1.66	73

such as, Rhamnazin-3-0-ß-D -glucopyranosyl-($l \rightarrow 5$)- α -Larabinofuranoside, Rhamnazin-3-O-ß-D- glucopyranosyl-($l \rightarrow 5$)-[ß-D-apiofuranosyl-($-1 \rightarrow 2$)]- α -L-arabinofuranoside reduced from 6.5 mg/L to 1.42 mg/L, from 6.5 mg/L to 1.66 mg/L, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (**Table 3**).

2-methoxy-5-methylaniline metabolite such as, 5-nitro-o-toluidine decreased from 134.6 mg/L to 36.34, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (Table 4). 2,4-diaminoanisole such as, 4-acetylamino-2-aminoanisole, 2,4-diacetylaminoanisole reduced from 275.8 mg/L to 22.06 mg/L, from 275.8 mg/L to 38.61 mg/L, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (Table 4). 4,40-diamino diphenyl ether metabolites such as, N,N¹-diacetyl-4,4¹-diaminobenzhydrol, N,N¹-diacetyl-4,4¹ diaminophenylmethane dropped off from 156 mg/L to 28.08 mg/L, from 156 mg/L to 40.56 mg/L, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (Table 4). o-aminoazotoluene metabolites such as, hydroxy-OAT (I), 4' -hydroxy-OAAT, 2' -hydroxymethyl-3-methyl-4-aminoazobenzene, 4, 4 -bis(otolylazo)-2, 2' -dimethylazoxybenzene decreased from 293.6 mg/L to 58.72 mg/L, from 293.6 mg/L to 79.27 mg/L, from 293.6 mg/L to 85.14 mg/L, from 293.6 mg/L to 117.44 mg/L, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L La-ZnO nanocomposite concentration at 50 W UV power and at 25°C (Table 4). 4-aminoazobenzol metabolites such as phenylhydroxylamine, nitrosobenzol reduced from 178 mg/L to 39.16 mg/L, from 178 mg/L to 44.5 mg/L, respectively, after 60 min photooxidation time, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power and at 25°C (Table 4).

[34] researched the photocatalytic activity of pure and Ni⁺²-doped BiO samples for the degradation of Rhodamine B (RhB). The effect of Ni⁺² doping concentration on the photocatalytic activity of RhB

Table 4: The metabolites of	pol	varomatic am	ines	in	the	тw
Table 4. The metabolites of	por	yaromatic am	mes		une	1

Deluganous stig	Daluanan atia aminaa	Influent	Effluent	Removal
Polyaromatic	Polyaromatic amines	concentrations	Concentrations	efficiencies
amines	metabonies	es Influent concentrations (co (mg/L) e 134.6 275.8 sole 275.8 sole 275.8 ¹ - ol 156 156 156 293.6 C 20 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2 2	(mg/L)	(%)
2-methoxy-5- methylaniline	5-nitro-o-toluidine	134.6	36.34	87
2,4-diaminoanisole	4-acetylamino-2- aminoanisole	275.8	22.06	92
	2,4-diacetylaminoanisole	275.8	38.61	86
4,40-diamino diphenyl ether	N,N ¹ -diacetyl-4,4 ¹ - diaminobenzhydrol	156	28.08	82
	N,N ¹ -diacetyl-4,4 ¹ -diaminophenylmethane	156	40.56	74
o-aminoazotoluene	hydroxy-OAT (I)	293.6	58.72	80
	4' -hydroxy-OAAT	293.6	79.27	73
	2' -hydroxymethyl- 3-methyl-4- aminoazobenzene	293.6	85.14	71
	4, 4'-bis(otolylazo)-2, 2' -dimethylazoxybenzene	293.6	117.44	60
4-aminoazobenzol	Phenylhydroxylamine	178	39.16	78
	Nitrosobenzol	178	44.5	75

was also investigated. 9 g/L Ni⁺²-doped BiO with a mass ratio of 2wt% had high photocatalytic efficiency [34]. 4-nitrophenol degradation was studied in the presence of Ni doped BiO nanoparticles with a Ni mass ratio of 4% [35]. 78.26% 4-nitrophenol removal was observed the aforementioned nanocomposite after 195 min photodegradation time and at 30 W UV light irradiation at pH=8 [35]. 68.57% Acid Yellow 29 55% Coomassie Brilliant Blue G250 and 37.27% Acid Green 25 degradations was obtained after 120 min of irradiation in the presence of 0.9% Ni-doped BiO, at 500 W UV light irradiation, under atmospheric oxygen, at 25°C, respectively [36]. The color and pollutant yields obtained in our study exhibited higher yields compared to the studies given above with low Ni-BiO nanocomposite concentrations.

Effect of increasing Ni mass ratios on 30 mg/L Ni doped BiO nanocomposite for photodegradation of TW pollutants

Were researched the effects of different La mass ratios (0.5wt%, 1wt%, 1.5wt% and 2wt%) in 30 mg/L Ni-BiO nanocomposite concentrations on the photooxidation yields of all pollutants in the TW during photooxidation experiments. Maximum $\text{COD}_{\text{total}}, \text{COD}_{\text{inert}}$ total flavonols, TAAs and color removal efficiencies were 99%, 92%, 91%, 98% and 99%, respectively, after 60 min photooxidation time, at pH=8.0, at 1.5wt% Ni mass ratio and at 25°C (Table 5 and Figure 7). Removal efficiencies increased as the Ni mass ratio in the Ni doped BiO nanocomposite were increased from 0.5wt% to 1wt% and to 1.5wt%. Maximum removal efficiencies was measured at 1.5wt% Ni mass ratio in the nanocomposite. The photocatalytic degradation efficiency of BiO nanoparticles increases with an increase in the Ni loading and shows a maximum activity at 1.5 wt%. Then decreases in photooxidation yield was observed on further Ni doping (to 2 wt%). The reason of this can be explained as follows: excessive amounts of dopants can retard the photocatalysis process, because excess amount of dopants deposited on the surface of BiO increases the recombination rate of free electrons

Table 5 Effect of increasing Ni mass ratios on the TW during photooxidation process after 60 min, at 50 W UV irradiation, 30 mg/L Ni-BiO nanocomposite concentrations, at pH=8.0, at 25°C.

	Removal efficiencies (%)					
	Ni mass ratios (%)					
Parameters	0.5wt%	1wt%	1.5wt%	2wt%		
COD _{total}	46	71	99	80		
COD	40	69	92	74		
COD	45	70	98	78		
Color	57	76	99	81		
Total flavonols	35	64	91	75		
Flavonols						
Kaempferol	30	63	87	68		
Quercetin	31	67	88	69		
Patuletin	32	68	90	75		
Rhamnetin	33	62	87	68		
Rhamnazin	30	60	85	67		
TAAs	53	81	98	77		
Polyaromatics						
2-methoxy-5-methylaniline	50	72	93	69		
2,4-diaminoanisole	49	77	95	75		
4,40-diamino diphenyl ether	47	64	87	64		
o-aminoazotoluene	44	71	84	70		
4-aminoazobenzol	41	70	82	66		



SCODtotal ■CODinert STotal Flavonols STAAs

Figure 7: Removal efficiencies of COD_{total} , COD_{inert} , total flavonols and TAAs at 0.5wt%, 1wt%, 1.5wt%, and 2 wt% Ni mass ratio.

and energized holes, thus inhibiting the photodegradation process. Hence, further increase in Ni doping to 2wt% results in the decrease of photocatalytic degradation efficiency.

The synthesized Ni-doped BiO catalyst possesses smaller particle size distribution than pure BiO nanoparticles. Apart from their small size, as Ni⁺² was doped in BiO, more surface defects are produced as reported by [37]. Consequently, the migration of the photo-induced electrons and holes toward surface defects is reasonable [37]. Thus, the separation efficiency of the electron–hole pairs of Ni-doped BiO with more oxygen defects should be more than that of the pure BiO nanoparticles. Therefore, the enhancement in the photocatalytic degradation efficiency of Ni doping BiO increases due to small particle size and higher defect concentration compared to BiO alone.

UV absorbances of Ni-doped BiO

The UV–vis absorption spectra of BiO and Ni-doped BiO are shown in **Figure 8**. It can be clearly seen from **Figure 8**. The maximum absorbance shifts is 410 nm for pure nano BiO while the maximum absorbance of Ni-doped BiO with a Ni mass ratio 0.5wt% is observed at a wavelentgh of 380 nm. The wave of absorbance of Ni-doped BiO also increases gradually with increasing the Ni loading and is much higher as compared to that of pure BiO. This could be mainly attributed to the quantum size effect as well as the strong interaction between the surface oxides of Bi and Ni. These observations strongly suggest that the Ni doping significantly affects the absorbance properties.

The strong UV band gap emission (375–395 nm) results from the radiative recombination of an excited electron in the conduction band with the valence band hole. The broad visible or deep-trap state emissions (410–440 nm and 540–580 nm) are commonly defined as the recombination of the electron-hole pair from localized states with energy levels deep in the band gap, resulting in lower energy emission. These deep-trap emissions indicate the presence of defects or oxygen vacancies of BiO nanostructures [38]. Since the band gap excitation of electrons in BiO or Ni-doped BiO with 254 nm can promote electrons to the conduction band with high kinetic energy, they can reach the



Figure 8: UV-vis absorption spectra of Ni doped BiO catalysts

solid-liquid interface easily, suppressing electron-hole recombination in comparison with 365 nm. Hence, the observation of low rate at 254 nm is therefore unexpected [39]. The UV band gap emission of Nidoped BiO nanostructures was increased between 380 and 410 nm after the photocatalytic process of pollutant parameters. The results show that 1.5 wt% Ni-doped BiO has maximum activity as compared to other photocatalysts.

Effect of increasing photooxidation time on the photooxidation yields of pollutants in the TW

Six different photooxidation times (5 min, 15 min, 30 min, 60 min, 80 min and 100 min) was examined during photocatalytic oxidation of the pollutants in the TW. To determine the optimum time for maximum removals these pollutant parameters in the TW. The maximum photocatalytic oxidation removals was observed at 60 min photooxidation time, at pH=8.0 using 30 mg/L Ni doped BiO with a La mass ratio of 1.5wt% at an UV power of 50 W (Figure 9). The removals of $\text{COD}_{\text{total}}$, $\text{COD}_{\text{inert}}$, total flavonols, total aromatic amines and color were found to increase linearly with increase in retention time from 5 min up to 80 min. A further increase in retention time to and 100 min lead to a decrease in yields of pollutant parameters. In other words the removal efficiencies of pollutant parameters (COD components, flavonols, polyaromatics, color) decreased for photooxidation time > 60 min since at long irradiation times since the surface energy of Ni doped - BiO decreases [40]. The photooxidation can form small molecules such as H₂O, carbonmonoxide (CO), CO₂ and benzene etc. after long irradiation; it will lead to the decrease of the polar groups and the oxygen content of pollutant surface. The dispersive component of surface energy, the density of polymer surface has great influence on dispersivity of pollutants in the TW. However, the rate of photodegradation of Ni doped-BiO blends increases with the increase of irradiation time, and is higher than that of photocrosslinking after long irradiation time, leading to the decrease of the density of the polymer surface and the dispersivity of COD, dyes and other pollutants to Ni doped-BiO [41]. The photooxidation can form small molecules such as H₂O, CO, CO₂ and benzene etc. after long irradiation; it will lead to the decrease of the polar groups and the oxygen content of



Figure 9: Removal efficiencies of COD_{total} , COD_{inter} , total flavonols and TAAs after 5, 15, 30, 60, 80 and 100 min retention times.

polymer surface, therefore the dispersivity decreaeses resulting in low photooxidation yields [41]. Aromatic and phenolic metabolites which would adsorb strongly onto titania surface and block significant part of photoreactive sites.

The maximum CODtotal, CODinert, total flavonols, total aromatic amines and color removal efficiencies were 99%, 92%, 91%, 98% and 99%, respectively, after 60 min photooxidation time, at 1.5 wt% Ni mass ratio in 30 mg/L Ni-BiO nanocomposite concentration, at pH=8.0 and at 25°C under 50 W irradiation (Table 6). Also, flavonols such as kaempferol, quercetin, patuletin, rhamnetin, rhamnazin removal efficiencies were 87%, 88%, 90%, 87% and 85%, respectively (Table 6). The photooxidation removals of polyaromatic amines such as, 2-methoxy-5-methylaniline, 2.4-diaminoanisole, 4.40-diamino diphenyl ether, o-aminoazotoluene, 4-aminoazobenzol were 93%, 95%, 87%, 84% and 82%, respectively, after 60 min at pH=8.0 and at 25°C (Table 6). Kaempferol, quercetin, patuletin, rhamnetin, rhamnazin concentrations decreased from 5.7 to 0.741 mg/L, from 9.2 to 1.104 mg/L, from 10.3 to 1.03 mg/L, from 7.2 to 0.936 mg/L, from 6.15 to 0.923 mg/L, respectively. 2-methoxy-5-methylaniline, 2.4-diaminoanisole, 4.40-diamino diphenyl ether, o-aminoazotoluene, 4-aminoazobenzol concentrations decreased from 134.6 to 9.422 mg/L, from 275.8 to 13.79 mg/L, from 156 to 5.46 mg/L, from 293.6 to 10.28, from 178 to 6.23 mg/L, respectively.

The color yields obtained in this study for TW are higher than the studies given below: [42] investigated the effects of Bi $_{0.95}$ Ni $_{0.05}$ O and Bi $_{0.90}$ Ni $_{0.10}$ O on the treatment of Methylene Blue (MB) dyestuff removal under 18 UV irradiation for 1 h. 81% color yields were observed for the aforementioned Ni-Bi-O nanocomposites, respectively [42]. [43] found 80% color yields based on Reactive Black 5 after 60 min irradiation time under 90 W irradiation using Bi-Ni nanocomposite.

Effect of increasing UV powers on the yields of pollutants in the TW

In this study, four UV light powers were used (10 W, 30 W, 50 W and 100 W) to detect the optimum UV irradiation power for

Fable 6 Effect of increasing photooxidation time on the TW during photooxidation process,
t 50 W UV irradiation, at pH=8.0, 30 mg/L Ni-BiO nanocomposite concentrations, 1.5
vt% Ni mass ratio, at 25°C.

	Removal efficiencies (%)					
Parameters	5 min	15 min	30 min	60 min	80 min	100 min
COD _{total}	56	69	87	99	99	99
COD	50	67	81	92	92	92
COD _{dissolved}	55	68	85	98	98	98
Color	67	74	88	99	99	99
Total flavonols	45	63	82	91	91	91
Flavonols						
Kaempferol	40	61	75	87	86	86
Quercetin	41	65	77	88	86	85
Patuletin	41	66	81	90	89	89
Rhamnetin	43	61	74	87	85	84
Rhamnazin	39	58	72	85	84	84
TAAs	63	78	84	98	98	98
Polyaromatics						
2-methoxy-5-methylaniline	60	71	86	93	93	92
2,4-diaminoanisole	59	75	82	95	94	94
4,40-diamino diphenyl ether	57	62	71	87	79	78
o-aminoazotoluene	55	69	78	84	82	80
4-aminoazobenzol	51	62	75	82	81	79

maximum photo-removal of the pollutant parameters in the TW using 30 mg/L Ni doped BiO nanocomposite with a Ni mass ratio of 1,5%w. The maximum photocatalytic oxidation removals was observed at 50 W UV light irradiation, at pH=8.0, after 30 min photooxidation time and at 25oC (Table 7 and Figure 10). The CODtotal, CODinert total flavonols, total aromatic amines and color were found to increase linearly with increase in UV light irradiation from 10 W, up to 30 W, up to 50 W, respectively (Table 7 and Figure 10). Further increase of UV power up to 100 W did not affect positively the pollutant yields. Maximum CODtotal, CODinert, total flavonols, TAAs and color

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	Removal efficiencies (%)						
Parameters		UV light i	rradiation				
	10 W	30 W	50 W	100 W			
COD _{total}	49	82	99	97			
COD _{inert}	43	76	92	90			
COD	48	81	98	96			
Color	60	83	99	99			
Total flavonols	38	76	91	90			
Flavonols							
Kaempferol	33	70	87	87			
Quercetin	34	71	88	86			
Patuletin	35	78	90	89			
Rhamnetin	36	70	87	87			
Rhamnazin	33	71	85	85			
TAAs	56	79	98	96			
Polyaromatics							
2-methoxy-5-methylaniline	53	81	93	92			
2,4-diaminoanisole	52	77	95	93			
4,40-diamino diphenyl ether	50	66	87	86			
o-aminoazotoluene	47	73	84	81			
4-aminoazobenzol	45	71	82	80			



№ 10 W ■ 30 W ∞ 50 W ≥ 100 W

Figure 10: Removal efficiencies of $\text{COD}_{\text{total}}$, $\text{COD}_{\text{inert}}$, total flavonols and TAAs at 10 W, 30 W, 50 W and at 100 W.

removal efficiencies after photooxidation process were 99%, 92%, 91%, 98% and 99%, respectively, for the aforementioned operational conditions (Figure 10). Flavonols such as kaempferol, quercetin, patuletin, rhamnatin, rhamnazin removal efficiencies were 87%, 88%, 90%, 87% and 85%, respectively, after 60 min photooxidation time, at 50 W UV light, at pH=8.0, at 30 mg/L Ni-BiO nanocomposite concentration and at 25oC(Table 7). Polyaromatic amines such as, 2-methoxy-5-methylaniline, 2, 4-diaminoanisole, 4, 40-diamino diphenyl ether, o-aminoazotoluene, 4-aminoazobenzol removal efficiencies after photooxidation process were 93%, 95%, 87%, 84% and 82%, respectively (Table 7).

The UV power determines the extent of light absorption by the semiconductor catalyst at a given wavelength. During initiation of photocatalysis, electron-hole formation in the photochemical reaction is strongly dependent on the optimum light intensity [44]. In this study, as the UV power increase from 10 W up to 50 W might favor a high-level surface defects, which account for the increase in the defect emission relative to the UV emission as reported by [39]. Higher UV powers > 50 W decrease the defects in the surface of the nanoparticle by disturbing the active holes.

Effect of increasing pH values on the pollutant yields in the TW

The effects of increasing pH values (4.0, 6.0, 8.0 and 10.0) on the photocatalytic oxidation of polutant parameters in TW was examined by considering the solubility of BiO nanoparticles in acidic as well as in highly basic solutions. The maximum photocatalytic oxidation removals was obtained at pH=8.0, after 60 min photooxidation time with a Ni mass ratio 1.5wt% using 30 mg/L Ni-BiO nanocomposite concentration at 50 W UV power (Table 8 and Figure 11). In acidic medium, less photocatalytic degradation of pollutant parameters (COD components, flavonols, polyaromatics, color) was observed. The extent of photocatalytic degradation of polutant parameters was found to increase with increase in initial pH to 8.0 and a decrease in maximum photocatalytic degradation was found at pH 10. The

possible explanation of this is that the pH at zero point charge (zpc) of BiO is 9.0 ± 0.3 [45]. Below pH 8.0, active sites on the positively charged catalyst surface are preferentially covered by pollutant molecules. Thus, surface concentration of the polutant parameters (COD components, flavonols, polyaromatics, color) is relatively high, while those of OH- and OH• are low. Hence, photocatalytic degradation decreases at acidic pH. On the other hand, above pH 8.0, catalyst surface is negatively charged by means of metal-bound OH-, consequently the surface concentration of the polutant parameters (COD components, flavonols, polyaromatics, color) is low, and OH• is high. In addition, polutant parameters are not protonated above pH 8.0. The electrostatic repulsion between the surface charges and Ni doped BiO nanocomposite hinders the amount of polutant parameters and the adsorption, consequently surface concentration of the polutant parameters and the adsorption, consequently surface concentration of the polutant parameters and the adsorption consequently surface concentration of the polutant parameters and the adsorption consequently surface concentration of the polutant parameters and the adsorption consequently surface concentration of the polutant parameters and the adsorption consequently surface concentration of the polutant parameters and the adsorption consequently surface concentration of the polutant parameters and the adsorption consequently surface concentration of the polutant parameters decreases, which results in the decrease

Table 8 Effect of increasing pH values on the TW during photooxidation process, at 50 WUV irradiation, after 60 min, at 25°C.

	Removal efficiencies (%)						
		pH val	ues				
Parameters	pH=4.0	pH=6.0	pH=8.0	pH=10.0			
COD _{total}	53	74	99	72			
COD	47	72	92	70			
COD	52	73	98	71			
Color	64	79	99	77			
Total flavonols	42	68	91	66			
Flavonols							
Kaempferol	37	66	87	64			
Quercetin	38	70	88	68			
Patuletin	39	71	90	69			
Rhamnetin	40	66	87	64			
Rhamnazin	45	62	85	60			
TAAs	60	83	98	81			
Polyaromatics							
2-methoxy-5-methylaniline	57	76	93	74			
2,4-diaminoanisole	56	80	95	78			
4,40-diamino diphenyl ether	54	67	87	65			
o-aminoazotoluene	52	74	84	72			
4-aminoazobenzol	50	65	82	63			



Figure 11: Removal efficiencies of COD_{total} , COD_{inert} , total flavonols and TAAs at pH=4.0, pH=8.0, pH=8.0, pH=10.0.

of photocatalytic degradation at pH 10.0. In conclusion, pH 8.0 can provide moderate surface concentration of polutant parameters which react with the holes to form $OH\bullet$.

Photocatalytic oxidation mechanisms of Ni doped BiO nanocomposite

The higher activity of Ni doped BiO can beattributed to successful e^--h^+ separation and production of ${}^{\bullet}O_2^-$ and OH ${}^{\bullet}$. Ni-modified BiO sample manifests the highest efficiency, which may be explained by the highest number of O_2 vacancies (related to the different charge and electronegativity of Ni and Bi ions) and as a result of stronger adsorption of OH $^-$ ions onto the BiO surface [46]. This favors the formation of OH $^{\bullet}$ by reaction of hole and OH $^-$. The OH $^{\bullet}$ and photogenerated ${}^{\bullet}O_2^-$ has extremely strong non-selective oxidants lead to the degradation of the organic pollutant at the surface of Ni modified BiO [41]. The photocatalytic degradation mechanism starts with the illumination of BiO nanoparticles and production of electron-hole pairs in Eq. (2):

$$BiO + h_{x}(UV) \rightarrow e^{-} + h^{+}$$
⁽²⁾

Major roles of metal ions in this study are to increase the concentration of BiO on the surface of the catalyst and to prolong the individual life-time of electrons and holes and hence, inhibit their recombination. The ability of Ni^{+3} to scavenge photogenerated electrons is as follows: (Eq. 3):

$$Ni^{+3} + e^{-} \rightarrow Ni^{+2} \tag{3}$$

However, stabilities of Ni^{+3} ions may be disturbed in their reduced forms (Ni^{+2}). This can be achieved by transferring the trapped electron to O_2 [Eq. (4)]:

$$\begin{array}{rrrr} \mathrm{Ni}^{+2} \ + \ \mathrm{O}_2 \ \rightarrow \ \mathrm{Ni}^{+3} \ + \ \mathrm{O}_2^{-\bullet} \\ (4) \end{array}$$

The produced $O_2^{-\bullet}$ is responsible from the generation of OH[•], known as highly reactive electrophilic oxidants [Eqs. (5-7)]:

$O_2^{-\bullet} + H^+ \rightarrow HO_2^{\bullet}$	(5)
$2HO_2^{\bullet} \rightarrow H_2O_2 + O_2$	(6)
$H_2O_2 + e^- \rightarrow OH^- + OH^{\bullet}$	(7)

In the meantime, photogenerated holes may react with H_2O molecules and produce OH^{\bullet} (Eq. 8):

$$h_{VB}^{+} + H_2 O \rightarrow H^+ + OH^{\bullet}$$

$$\tag{8}$$

The color removal by photooxidation of dyes reactions were given in Eqs (9-14):

aromatics/phenol + $h_v \rightarrow Dye^*$ (9)

Aromatics/phenol* +	Ni doped BiO	\rightarrow	Aromatics/phenol ⁺	+	
Ni - BiO(e)					(10)

$$Ni - BiO(e) + O_2 \rightarrow NiO - BiO + O_2 \tag{11}$$

$$O_2^{-\bullet} + Ni - BiO(e) \rightarrow 2H^+ + H_2O_2 \tag{12}$$

$$H_2O_2 + Ni - BiO(e) \rightarrow OH^{\bullet} + OH^{-}$$
(13)

 $Dye^{+} + O_{2} (or O_{2}^{-\bullet} or OH^{\bullet}) \rightarrow$ $peroxylated or hydroxylated intermediates \rightarrow$ degraded or mineralized products(14)

Thus, loading of metal ions such as Ni on the surface of BiO matrix can suppress the recombination of photoinduced charge carriers either with only electron capture ability or with steps forward to produce OH[•]. For Ni–BiO, electron accepting ability, production of more OH[•], the highest surface roughness value and the higher dark adsorption capacity result in pronounced photoactivity. The decay profile of the products includes the subsequent attacks of OH[•], known as highly reactive electrophilic oxidants. The main reaction pathway (60% of OH[•]) is the addition of the OH[•] to the double bond of the azo group, resulting in the rapid disappearance of color; however, addition to the aromatic ring also occurs (40% of OH[•]) [47, 48]. Further OH[•] attacks and the increment in OH[•] concentration in the solution increase the yield of OH-adduct in the degradation progress of each product. The opening of the dye aromatic rings due to consecutive oxidation reactions leads to low-molecular weight compounds [49].

Photonic efficiency of Ni doped BiO

In order to evaluate the relative photonic efficiency (I) a solution of MCP (40 mg/L) adjusted to pH 10 was irradiated with 100 mg BiO (Merck) and Ni-doped BiO, separately. The relative photonic efficiencies of light of wavelengths 254 and 365 nm for BiO and Nidoped BiO are presented in Table 9. For comparison, the relative photonic efficiency of TiO, is also presented in Table 9. The relative photonic efficiencies of Ni-doped BiO are greater as compared to those of BiO and TiO,, revealing the effectiveness of metal-doped systems. It is also interesting to note that the relative photonic efficiency for Ni-doped BiO for light of wavelength 254 nm are much higher as compared to that for 365 nm. The results are in good agreement with degradation and mineralization studies. Comparing the high efficiency of Ni doped BiO catalysts with standard BiO and TiO₂ catalyst, the photocatalytic efficiency of 1.5 wt% Ni-doped BiO is higher as compared to that of BiO and TiO, and other Ni doped BiO nanacomposites.

Reusability of Ni doped BiO

As shown in **Figure 12**, after the first cycle of photocatalytic oxidation within 60 min, 99% of the Ni doped BiO with a mass ratio of 1.5wt% was recovered. After three cycles, the phoooxidation ability of Ni doped BiO nanocomposite was retained at 93% of the original value. After 8th cycles the nanocomposite was reatined at 80%. One of the reasons for the slight decline in photooxidation is that the surface of the reused photocatalysts may exist with some low residual substances which did not occupy the photocatalytic sites and did not block the adsorption. The presence of Ni significantly changed the binding

Parameters	Relative photonic efficiency (I)	
	256 nm	370 nm
Pure BiO	1.01 ± 0.001	0.79 ± 0.01
0.1wt% Ni-BiO	1.09 ± 0.01	0.93 ± 0.01
0.5 wt% Ni-BiO	1.28 ± 0.02	0.94 ± 0.01
1.0wt% Ni-BiO	2.59 ± 0.01	2.22 ± 0.01
1.5wt% Ni-BiO	2.98 ± 0.01	2.25 ± 0.01
2wt% Ni-BiO	1.02 ± 0.01	1.98 ± 0.01
Commercial BiO	1.02 ± 0.01	1.02 ± 0.01
TiO ₂	1.38 ± 0.01	1.03 ± 0.01

(*): pH 5; UV = 8 lamps; \ddot{e} = 254 and 365 nm; 50 W UV power, 60 min photooxidation time

Table 9 Comparison of relative photonic efficiencies in the photodegradation of pollutants in TW by BiO and Ni-doped BiO photocatalysts (*)



site of the pollutant molecules. It is possible that the oxygen atom in Ni-BiO was bound to the dopant Ni [50]. The speedily recovering of the photodegradation capacity of Ni doped BiO for pollutans photodegradation will benefit to their photocatalytic activity.

Conclusions

By using 30 mg/L Ni-BiO with a Ni mass ratio of 1.5w% the CODtotal, CODinert, flavonols, polyaromatics and color were photodegraded with yields as high as 82-99% within 60 min photooxidation time, at 25oC under 50 W UV power, at pH=8.0. The addition of Ni to BiO lead to enhance the photocatalytic activity by increasing the total surface area. The flavonoids and polyaromatic amines and their metabolites in the TW were firstly determined photodegraded with high rates and photonic efficiency using 30 mg/L Ni-BiO with a Ni mass ratio of 1.5w% at pH 8.

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